



1 Topsoil organic carbon storage of China and its loss 2 by cultivation

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7 **Key words:** Carbon pool, China soils, Cultivation-induced change, Global change, SOC, Topsoil

8 **Abstract.** Topsoil is very sensitive to human disturbance under the changing climate. Estimates
9 of topsoil soil organic carbon (SOC) pool may be crucial for understanding soil C dynamics
10 under human land uses and soil potential of mitigating the increasing atmospheric CO₂ by soil
11 C sequestration. China is a country with long history of cultivation. In this paper, we present
12 an estimate of topsoil SOC pool and cultivation-induced pool reduction of China soils based
13 upon the data of all the soil types identified in the 2nd national soil survey conducted during
14 1979–1982. The area of cultivated soils of China amounted to 138×10^6 ha while the unculti-
15 vated soils occupied 740×10^6 ha in 1980. Topsoil SOC density ranged from 0.77 to
16 1489 t Cha⁻¹ in uncultivated soils and 3.52 to 591 t Cha⁻¹ in cultivated soils with the average
17 being 50 ± 47 t Cha⁻¹ and 35 ± 32 t Cha⁻¹, respectively. Geographically, the maximum mean
18 topsoil SOC density was found in northeastern China, being of 70 ± 104 t Cha⁻¹ for unculti-
19 vated soils and of 57 ± 54 t Cha⁻¹ for cultivated soils, respectively. The lowest topsoil SOC
20 density for uncultivated soils was found in East China, being of 38 ± 33 t Cha⁻¹ and that for
21 cultivated soils in North China, being of 30 ± 30 t Cha⁻¹. There is still uncertainty in esti-
22 mating the total topsoil SOC of uncultivated soils because a large portion of them was not
23 surveyed during the 2nd Soil Survey. However, an estimate of total SOC for cultivated soils
24 amounted to 5.1 Pg. On average, cultivation of China's soils had induced a decrease of SOC
25 density of 15 t Cha⁻¹ giving rise to an overall pool reduction at 2 Pg. This is significantly
26 smaller than the total SOC pool decline of 7 Pg due to cultivation of natural soils in China
27 reported by Wu et al. (*Glob. Change Biol.* 2003, 9: 305–315), who made a pool estimation of
28 whole soil profile assuming 1 m depth for all soils. As the mean topsoil SOC density of China
29 was lower than the world average value given by Batjes (*J. Soil Sci.* 1996, 47: 151–163), China
30 may be considered as a country with low SOC density and may have great potential for C
31 sequestration under well defined management. However, the dynamics of topsoil C storage in
32 China agricultural soils since 1980's and the effects of modern agricultural developments on C
33 dynamics need further study for elucidating the role of China agriculture in global climatic
34 change.

35

36 Introduction

37 Soil organic carbon (SOC) storage in terrestrial C cycling under global cli-
38 mate change has become one of the foci of global soil studies (Lal 1999;
39 Schlesinger 1999, 2000; Kirschbaum 2000; Amundson 2001; Rustad et al.
40 2001). Estimating the carbon pool and potential sink effect of soils may be

41 crucial for a country or a region to commit to the Kyoto Protocol and
42 Global Climate Change Framework Agreement (Smith et al. 2000a, b) as
43 carbon sequestration by soils has been considered as a practical measure for
44 mitigating the rise in atmospheric CO₂ (Lal 1999; Schlesinger 1999, 2000).
45 World soils preserved approximately 1500 Pg SOC in the upper 1 m of soil
46 cover (Batjes 1996; Lal 1999), changes in climate and land use may have
47 significant effects on SOC dynamics, particularly with respect to its turnover
48 rate (Rustad et al. 2001; West and Marland 2002). Changes in SOC of
49 agricultural soils have been reported by Eve et al. (2002). Assessment of pool
50 size and turnover of SOC at different scales (Batjes 1996, 2002; Fearnside and
51 Barbosa 1998; Batjes and Dijkshoorn 1999; Houghton et al. 1999; Bhat-
52 tacharya et al. 2000; Bhatti et al. 2002a, b; Vleeshouwers and Verhagen 2002;
53 West and Marland 2002) have been very well documented while addressing
54 the role of soil carbon dynamics in global change. However, there have been
55 few studies of total topsoil SOC pools of cultivated soils at national or global
56 level.

57 Efforts have been made worldwide to enhance soil C sequestration to
58 offset CO₂ emission from the industrial sector (Lal et al. 1999; Schlesinger
59 1999, 2000; Lal 2002a). Increasing attention had been paid to SOC
60 dynamics and its potential for C sequestration in croplands for the last
61 decade (Dadal and Mayer 1986; Rounsevell et al. 1999; Smith et al. 2000a,
62 b; Jacinthe et al. 2001; Uri 2001; Hao et al. 2002; Schuman et al. 2002).
63 Such issues have been raised especially for China with its intensely culti-
64 vated soils under extensive soil degradation in a process of fast industrial-
65 ization (Lal 2002b). China is a country with a long history of cultivation of
66 diverse soil types. While the industrial C emission of China has been rising,
67 C loss from China soils due to intensive agricultural land use has also raised
68 serious concerns (Lindert et al. 1996; Li 2000; Lal 2002b; Wu et al. 2003).
69 Various estimates of the total SOC pool of China's soils range from 50 to
70 200 Pg (Wang and Zhou 1999; Ni 2001; Pan et al. 2003b). However, the
71 topsoil SOC pool size and its dynamics have been poorly studied (Pan et al.
72 2003b). Supposing that topsoil SOC stock may account for 80–90% of the
73 stock variations to be observed over decades, Arrouys and Balesdent (2002)
74 worked out an estimate of topsoil SOC pool size of French soils to the
75 upper 30 cm by using 19,000 references available in a national database.
76 Pan et al. (2003a) accomplished an estimate of topsoil SOC pool of China's
77 paddy soils by using the sampling depth records of the topsoil thickness.
78 Nevertheless, there had been no data on the SOC pool of croplands of
79 China and its change due to cultivation.

80 In this paper, we analyzed the data of all the China's soil types and esti-
81 mate the total topsoil SOC pool in an attempt to present the pool size and its
82 cultivation-induced change. We aimed to address the role of agriculture in
83 SOC storage dynamics and the necessity to approach a carbon sequestration
84 strategy for China's sustainable agriculture in mitigating the atmospheric
85 CO₂ rise.

86 **Data and methods**87 *Data source*

88 All the data was obtained by the 2nd State Soil Survey conducted in 1979–
 89 1982, which are available in a series of China Soil Types of Volumes 1–6
 90 (SSSSC 1993; 1994a, b; 1995a, b; 1996a, 1996b; 1997; 1998). The original soil
 91 data of the overall 2456 soil types identified by the soil survey was grouped into
 92 uncultivated and cultivated soils according to the sampling records. Soil
 93 sampling was done generally at a scale of 1:200 ha and locations were shown in
 94 Figure 1. As required by the 2nd State Soil Survey, the SOC content, thickness
 95 and bulk density of topsoil of typical uncultivated soil profiles and of most
 96 heavily cultivated soils were determined. As a whole, the data set comprised
 97 34,411 whole soil profiles and 523,894 topsoil samples. In this data set, 2553
 98 soil profiles were documented with their land use conditions, of which 923 soil
 99 profiles and 165,122 soil samples from uncultivated and 1630 soil profiles and
 100 358,772 soil samples from cultivated soils, respectively. The numbers of soil
 101 area were respectively 740×10^6 ha and 138×10^6 ha for uncultivated and
 102 cultivated soils though a large portion of uncultivated soils was not surveyed.

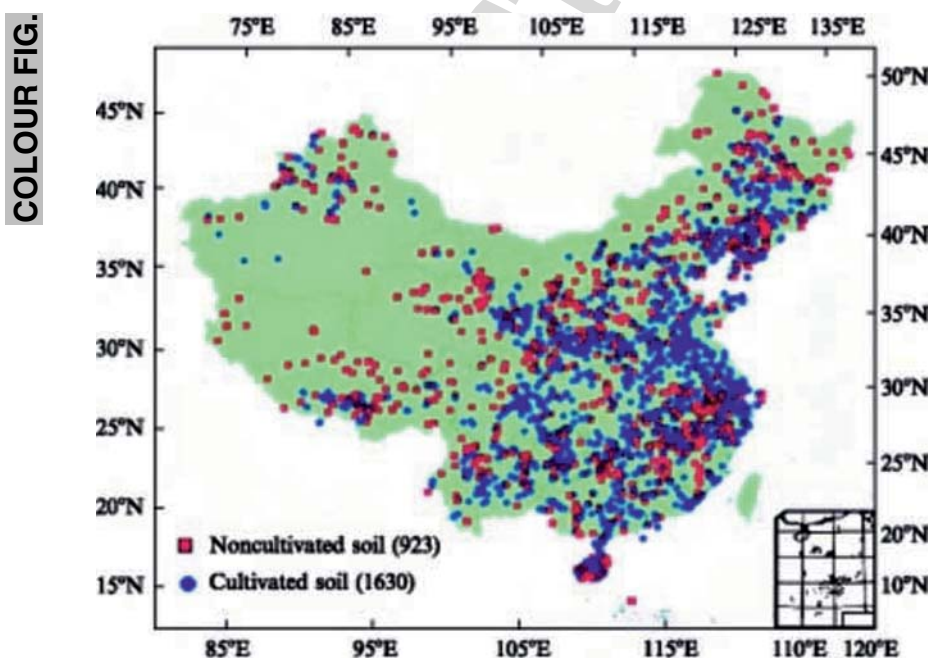


Figure 1. ■Au: Please provide Figure 1 caption.■

103 *Calculation of SOC pool estimate*

104 The SOC pool was calculated to the recorded depth of A horizon (or Ap and P
105 horizon in case of paddy soils) or to a depth of 30 cm in case of A horizon
106 thickness exceeding 30 cm. The SOM content from the original data was
107 converted to SOC by multiplying a constant of 0.580, since the determination
108 was done by conventional wet combustion (SSSSC 1996b). Thus, the topsoil
109 SOC density can be obtained by the following equation:

$$D_{oc} = SOC \times \gamma \times H \times (1 - \delta_{2\text{mm}}/100) \times 10^{-1} \quad (1)$$

111 where D_{oc} and SOC are the amount (t ha^{-1}) and content (g kg^{-1}) of SOC,
112 respectively, γ is the bulk density (g cm^{-3}), H is the recorded thickness (cm),
113 and $\delta_{2\text{mm}}$ is the fraction (%) of >2-mm fragments in soil. In cases where data
114 were missing, bulk density value γ in the equation was estimated by regression
115 analysis between the available data of bulk density and SOC content for a
116 given layer (Figures 2 a and b).

117 While the total SOC pool (P_{oc}) of soils can be estimated by:

$$P_{oc}(tC) = \sum_{i=1}^n S_i \times \sum_{j=1}^n SOC_j \times \gamma_j \times H_j \times 10^{-1} \quad (2)$$

119 where j is the sublayer number of topsoil and S_i is the number of area (ha) of a
120 given soil types i . The calculation was carried out separately for the cultivated
121 and uncultivated soils. The cultivation-induced C change was deduced by
122 subtracting the SOC amount of uncultivated soil by that of cultivated soil for
123 individual soil types.

124 **Results and discussion**

125 *Relationship between SOC content and bulk density*

126 The regressions between soil bulk density and SOC content depends on soil
127 types (Callsen et al. 2003; Pan et al. 2003a). Of the data available, 3645 topsoil
128 samples of uncultivated soils and 4765 of cultivated ones had records of means
129 or single measurements of both SOC content and bulk density (γ). The
130 regression between γ and SOC (Figures 2 a and b) are found as follows: For the
131 uncultivated soils:

$$\gamma = 1.3565 \times e^{-0.0046 \cdot \text{SOC}} \quad (R^2 = 0.7260, p < 0.001) \quad (3)$$

133 and for the cultivated soils

$$\gamma = 1.3770 \times e^{-0.0048 \cdot \text{SOC}} \quad (R^2 = 0.7870, p < 0.001) \quad (4)$$

135 The regression Eqs. (3) and (4) were used to estimate the missing bulk density
136 values for the uncultivated and cultivated soils, respectively.

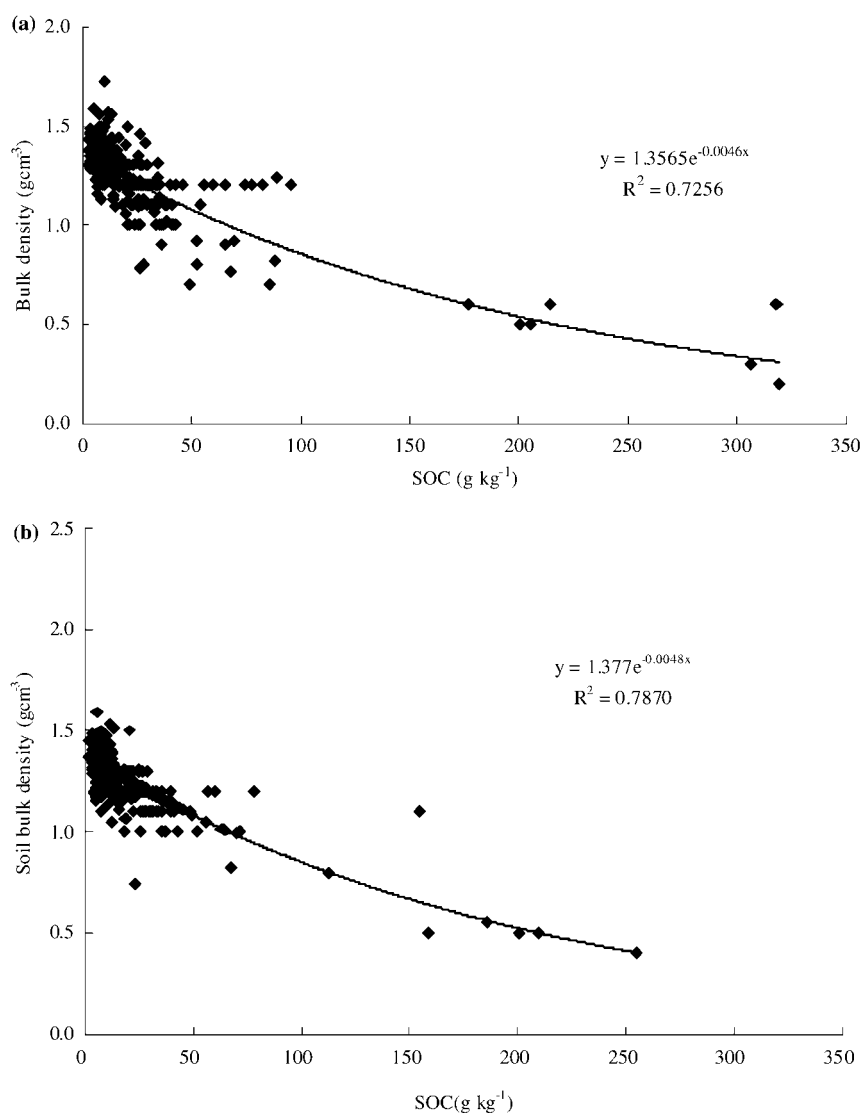


Figure 2. Correlation of bulk density with SOC for uncultivated soils (a) and for cultivated soils (b).

137 Topsoil SOC amounts

138 The calculated topsoil SOC density for individual soil types varied in a wide
 139 range from 0.77 to 1489 t Cha⁻¹ for uncultivated soils and from 3.52 to
 140 591 t Cha⁻¹ for cultivated soils. The frequency distribution patterns of SOC of

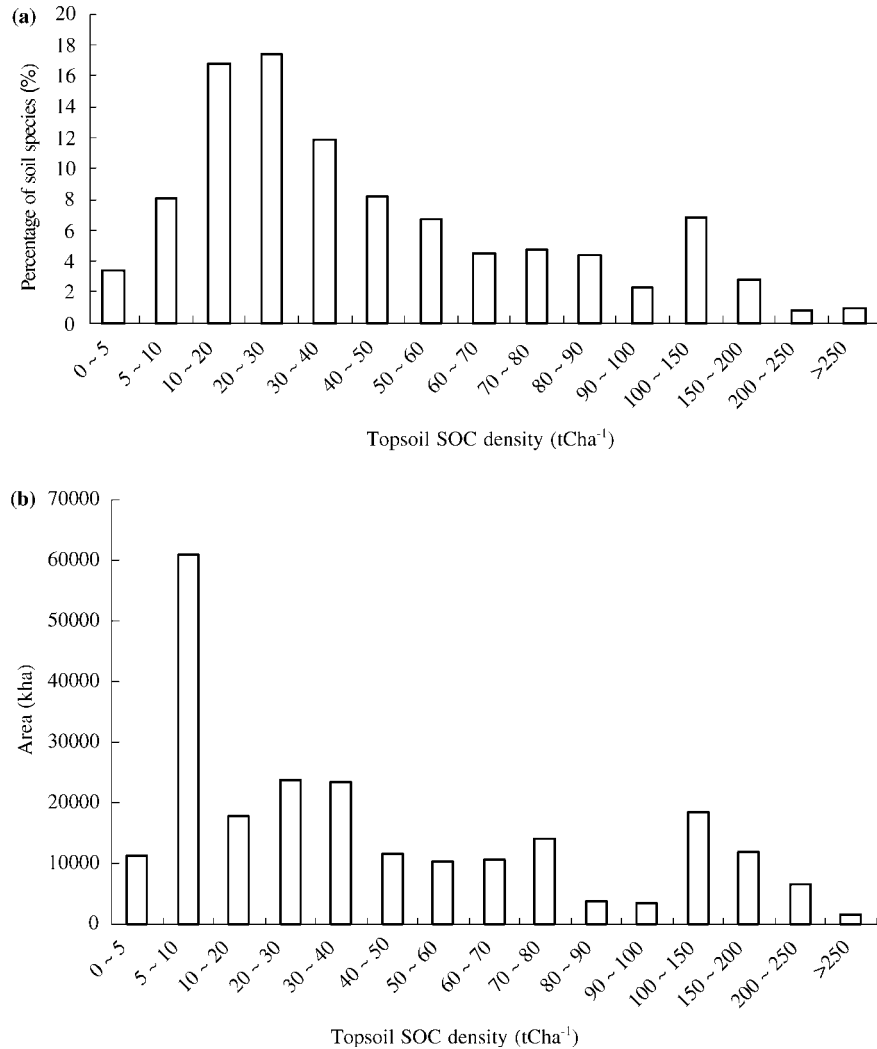


Figure 3. Frequency distribution of topsoil SOC density in the term of (a) percentage of number soil species (%) and (b) area (kha) among uncultivated soils.

141 1281 uncultivated soils and that of 1383 cultivated soils are shown in Figures 3
 142 and 4, respectively. Over 60% of the cultivated soils corresponding to an area
 143 of 82 Mha possessed an averaged SOC amount in range of 10 to 40 t Cha⁻¹,
 144 while only a minor portion of uncultivated soils (an area of 4.07 Mha) showed
 145 high averaged topsoil SOC in range of 100 to 200 t Cha⁻¹. The mean topsoil
 146 SOC density (Table 1) of uncultivated soils was 50 ± 47 t Cha⁻¹ and that of
 147 cultivated ones was 35 ± 32 t Cha⁻¹, showing a general reduction of topsoil

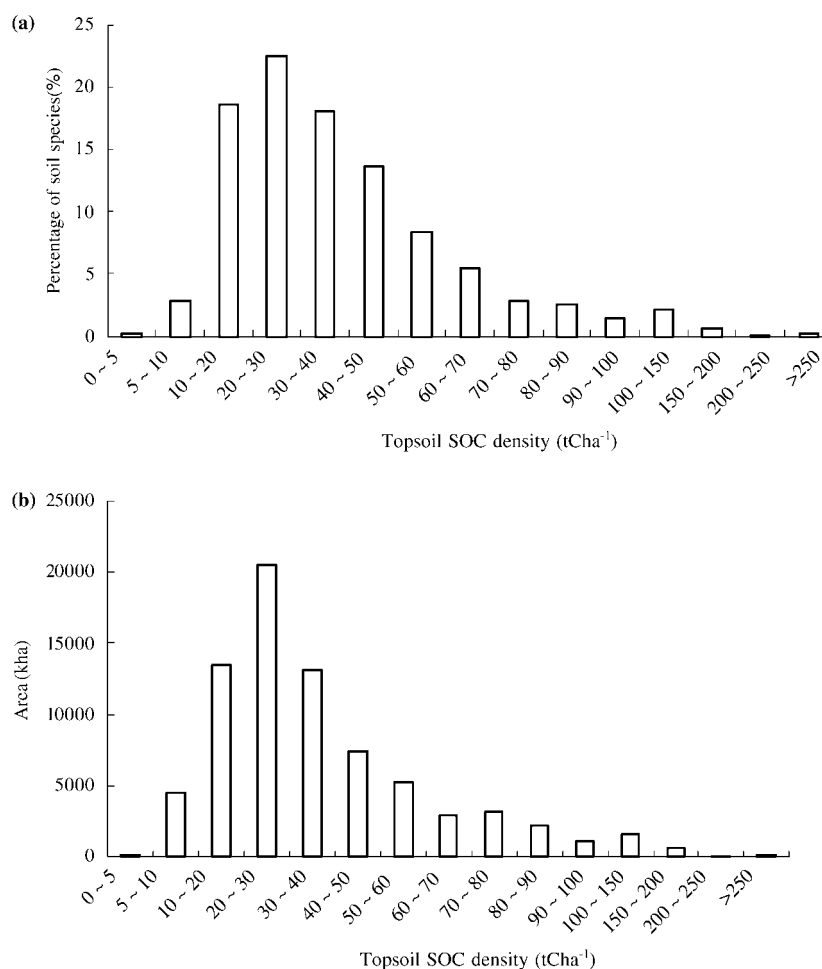


Figure 4. Frequency distribution of topsoil SOSOC density in the term of percentage of total number of soil species (a) and of occupying area (kha) (b) among cultivated soils.

148 SOC density at 15 t Cha⁻¹ on mean due to cultivation of natural soils. Simi-
 149 larly, Arrouys and Balesdent (2002) reported that mean SOC of French soils
 150 under different land uses ranged from 30 to 90 t Cha⁻¹ while of those under
 151 annual crops and perennial crops it was lower than 45 t Cha⁻¹ and those under
 152 permanent grassland and forests exhibited higher SOC density up to
 153 70 t Cha⁻¹. Their study showed a marked reduction in topsoil SOC pool after
 154 shifting from grassland and forests lands to farmlands. In Brazil where the
 155 SOC ranged from 15 to 418 t Cha⁻¹, most of the soil areas had an SOC
 156 varying between 30 and 60 t Cha⁻¹ (Bernoux 2002). Comparatively, China's
 157 soils generally had lower topsoil SOC density.

Table 1. Distribution of topsoil C storage among the groups of uncultivated and cultivated soils in China (data source: SSSSC 1997).

FAO/UNESCO	Area (kha)		Sample number		SOC density (t Cha ⁻¹)		C pool (Tg) ^a	
	Uncultivated	Cultivated	Uncultivated	Cultivated	Uncultivated	Cultivated	Uncultivated	Cultivated
Acrisols	2856.13	1074.00	424	606	28.91 ± 12.07	33.36 ± 16.49	86.08	33.35
Alisols	20131.8	3115.53	2656	1340	61.87 ± 33.14	37.50 ± 20.68	1259.24	126.15
Arenosols	112396.89	1066.40	1398	4934	20.22 ± 31.30	25.11 ± 33.50	968.86	17.04
Calcisols	61065.07	3288.40	2353	4686	14.90 ± 15.90	26.94 ± 12.63	451.88	93.29
Cambisols	185721.75	16539.90	32,287	40,898	45.70 ± 42.57	24.57 ± 11.82	14846.60	378.27
Chernozems	9234.60	3976.00	19,660	19,951	76.49 ± 54.60	71.59 ± 48.71	908.50	255.18
Fluvisols	6404.67	25989.53	14,070	49,187	18.65 ± 10.17	24.72 ± 14.84	138.92	553.84
Gleysols	12099.00	507.73	564	288	166.43 ± 252.21	121.92 ± 104.52	1844.13	84.07
Histosols	1442.27	38.93	84	66	246.95 ± 160.30	244.28 ± 198.40	186.20	16.54
Kastanozems	35195.67	7109.40	5589	6335	40.58 ± 23.21	36.28 ± 18.86	1488.07	258.00
Lixisols	930.60	905.60	868	12,751	24.31 ± 10.89	20.35 ± 7.09	26.20	18.14
Luvissols	28618.4	4014.33	9967	9571	74.77 ± 50.76	52.71 ± 26.08	1709.29	169.04
Phaeozems	2523.67	4822.87	15,423	3748	80.68 ± 30.75	72.85 ± 33.44	222.28	368.80
Podzols	0.067	-	1	-	133.62 ± 0.00	-	0.01	0.00
Regosols	20758.53	10411.93	16,028	16,574	26.99 ± 19.79	20.04 ± 9.83	538.06	164.09
Solonchaks	41243.14	717.93	2422	1572	18.27 ± 14.00	21.75 ± 14.61	483.37	16.20
Acrisols/Alisols	70486.13	4202.67	13,136	4530	30.95 ± 16.75	33.44 ± 17.87	2467.72	142.89
Fluvisols/Cambisols	-	29780.33	-	150,543	-	46.91 ± 25.73	0.00	1309.74
Gleysols/Phaeozem	18327.60	6742.40	9115	5656	65.25 ± 35.69	56.82 ± 29.75	1472.07	507.03
Leptisols/Cambisols	4103.67	78.67	144	16	151.68 ± 67.02	83.84 ± 44.26	569.38	4.47
Luvissols/Cambisols	54556.66	5785.27	8256	8128	73.72 ± 45.98	51.80 ± 35.72	6754.95	300.62
Regosols/Leptisols	51534.27	3869.86	8825	7312	42.79 ± 33.63	39.05 ± 24.66	2026.33	178.75
Vertisols/Cambisols	84.40	3676.67	1852	10,080	34.97 ± 11.72	28.90 ± 5.79	2.67	96.33
Total/Mean	739714.99	137714.35	165,122	358,772	49.84 ± 46.69(51.98) ^b	35.08 ± 31.57(36.97)	38450.82	5091.83

^aEstimated using the area weighted mean SOC density.^bThe number in parenthesis is the area weighted mean of all the soil groups.

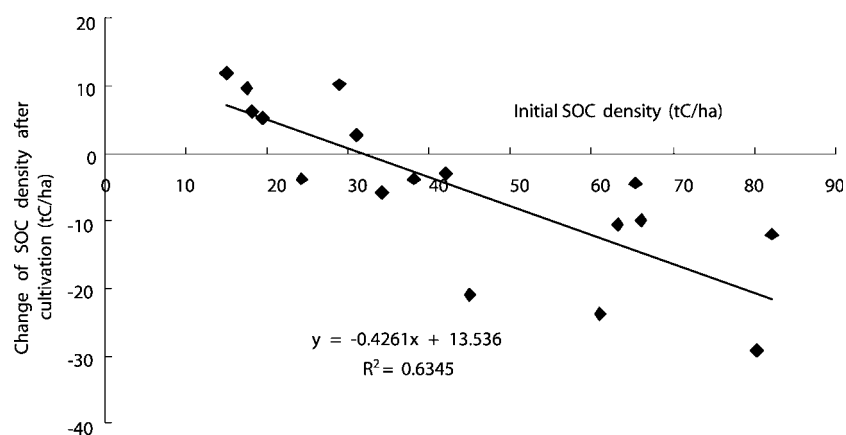


Figure 5. Change of mean topsoil SOC density of mineral soil groups after cultivation.

158 *Distribution of SOC amount in pedogenic soil groups*

159 The area number, means of SOC amounts and the calculated C pool of ped-
 160 ogenic soil groups in accordance to the system by FAO & UNESCO (1988)
 161 soils were given in Table 1. For uncultivated soils, the highest mean SOC
 162 density was found in Histosols being $247 \pm 160 \text{ t Cha}^{-1}$, and the lowest one in
 163 Calcisols being $15 \pm 16 \text{ t Cha}^{-1}$, of which most were in Northwest China
 164 where dry condition reduces crop growth depressing SOC accumulation. For
 165 cultivated soils, however, the highest mean SOC density was also found in
 166 Histosols being $244 \pm 198 \text{ t Cha}^{-1}$, the lowest one was in Regosols being
 167 $20 \pm 10 \text{ t Cha}^{-1}$. Cambisols had the biggest topsoil C pool for uncultivated
 168 soils at 15 Pg with a medium SOC density of $46 \pm 43 \text{ t Cha}^{-1}$ and the largest
 169 area of 186 Mha. The transitional group of Fluvisols/Cambisols was shown as
 170 a soil type of the biggest topsoil C pool (1.31 Pg) among the cultivated soils
 171 and had relatively high SOC amount of $47 \pm 26 \text{ t Cha}^{-1}$. In fact, most of
 172 these soils were classified as paddy soils in Chinese Pedogenic Classification in
 173 (SSSSC 1998). Pan et al. (2003a) had reported an enhanced C storage in paddy
 174 soils of China of 0.3 Pg due to long term paddy management. The effect of
 175 cultivation on topsoil SOC differed from group to group, with those high in
 176 SOC contents susceptible to change promptly (Figure 5). Cultivation did en-
 177 hanced SOC accumulation in some soil groups originally poor in SOM due to
 178 application of organic manure and/or high biomass input under irrigation (Pan
 179 et al. 2003b), especially in soil groups of dry farmlands of West China (Wang et
 180 al. 1996; 2001). This is also true for some soil groups from South China. Batjes
 181 (1996) reported worldwide mean topsoil (0–30 cm) SOC density values of 31,
 182 13, and 51 t Cha^{-1} of Acrisols, Arenosols, and Luvisols, respectively.
 183 Apparently, the topsoil SOC of cultivated soils of Acrisols, Arenosols and
 184 Luvisols in South China was comparable to or slightly higher than the world

185 means. Enhancement of SOC in cultivated Alisols and Arenosols in subtropical
186 China were frequently reported (Department of Agriculture, Hunan Province
187 1989; Liu et al. 1999).

188 *Distribution of SOC by geographical regions*

189 The distribution of topsoil SOC in terms of soil–geographical regions is shown
190 in Table 2. The highest mean topsoil SOC both for uncultivated and cultivated
191 soils was found in northeastern China, where high SOC amounts were fre-
192 quently reported (Wang et al. 2002; Li et al. 2004). The mean SOC amounts of
193 cultivated soils was $57 \pm 54 \text{ t Cha}^{-1}$ compared to $70 \pm 104 \text{ t Cha}^{-1}$ of
194 uncultivated soils in this region. The lowest mean topsoil SOC of uncultivated
195 soils ($38 \pm 33 \text{ t Cha}^{-1}$) was found in East China, the lowest of the cultivated
196 soils ($30 \pm 30 \text{ t Cha}^{-1}$) was observed in North China. Considerable reduction
197 of topsoil SOC in range of $20\text{--}40 \text{ t Cha}^{-1}$ was found in Southwest China,
198 Northeast China and North China, which are generally considered zones
199 vulnerable to degradation (Zheng et al. 1997; Zhou 1999). The biggest reduc-
200 tion of SOC density in cultivated soils of Southwest China could be attributed
201 to the severe desertification of the Karst lands due to improper cultivation of
202 the sloping and stony lime stone terrains (Anonymous 2003). Nevertheless,
203 there had been small reductions of SOC due to cultivation in South China. The
204 natural upland and Savanna soils in this region had SOC density values below
205 110 t Cha^{-1} for 1 m soil. While the paddy soils, as the main soils under cul-
206 tivation in this region, had a mean SOC density close to that of the soils under
207 needle-leaf forest and higher than that of dry upland farmlands (Zhao et al.
208 1997; Li and Zhao 2001). While in many cases, increase of SOC content in
209 cultivated soils was found as most of the red soils are poor in SOC in the
210 region.

211 Wang et al. (2002) and Li et al. (2004) reported a wide range of SOC of
212 $24\text{--}925 \text{ t Cha}^{-1}$ with an area-weighted mean of over 200 t Cha^{-1} for the upper
213 1 m in Northeast China under various vegetation types. The dramatic reduc-
214 tion of the topsoil SOC density due to cultivation and the wide range of SOC in
215 this region may reflect a high sensitivity of soil carbon in the temperate zone to
216 management (Wang et al. 2002). Many studies have discussed the sensitivity of
217 the ecosystems in high latitude regions to global climatic change (Bousquest
218 et al. 1999), and the soils in such regions are especially susceptible to future
219 land use change and projected climate change (Esser 1987; Tian et al. 2000). In
220 fact, Xie (1999) had pointed out the high sensitivity of the terrestrial ecosystem
221 of Northeast China to global warming. The SOC also decreased due to culti-
222 vation in northwestern China despite the very low mean SOC density
223 ($42 \pm 49 \text{ t Cha}^{-1}$ and $33 \pm 23 \text{ t Cha}^{-1}$, respectively for uncultivated and
224 cultivated soils). In this region, depletion of SOC in natural soils could be
225 attributed to unfavorable plant growth (Wang et al. 2001) and enhanced
226 decomposition and mineralization of biomass caused by aeration under

Table 2. Distribution of density and storage of topsoil SOC in geographical regions of China.

Geographical region	Surveyed area (Mha)		Topsoil thickness (cm)		SOC density (t Cha ⁻¹)		C storage (Tg)		Cultivation-induced loss	
	Uncultivated soils	Cultivated soils	Uncultivated soils	Cultivated soils	Uncultivated soils	Cultivated soils	Uncultivated soils	Cultivated soils	SOC density ^b (t Cha ⁻¹)	Pool (Tg)
East China	45.93	37.82	17.77 ± 7.00	23.46 ± 5.51	37.75 ± 32.83 (39.26) ^a	35.62 ± 16.81 (34.21)	1803.21	1293.82	5.05	191.0
Northeast China	55.52	21.17	25.56 ± 9.13	26.12 ± 7.40	70.49 ± 104.46 (100.53)	57.17 ± 53.67 (64.00)	5581.43	1354.88	36.53	773.3
South China	22.85	7.73	19.68 ± 6.52	22.77 ± 5.14	43.68 ± 32.64 (41.44)	38.02 ± 18.50 (39.47)	946.90	305.10	1.97	15.23
North China	17.53	18.06	24.98 ± 8.78	25.63 ± 7.38	39.96 ± 39.91 (47.64)	30.35 ± 29.93 (23.56)	835.13	425.49	24.08	434.9
Northwest China	221.97	25.21	23.32 ± 9.30	26.74 ± 4.73	41.74 ± 49.44 (36.73)	32.67 ± 22.76 (24.29)	8152.96	612.35	12.08	304.5
Southwest China	68.98	19.50	21.14 ± 8.33	24.49 ± 6.07	74.89 ± 81.74 (75.05)	48.29 ± 24.58 (42.08)	5176.95	820.56	32.97	642.9
Total/Mean	432.78	129.49	22.73 ± 8.88	24.81 ± 6.36	51.02 ± 66.11 (51.80)	38.41 ± 31.15 (37.16)	22416.67 (35249.41) ^c	4811.85 (5117.30) ^c	18.24	2.36

^aThe number in parenthesis is the area weighted mean.

^bArea weighted mean of different types of cultivated soils in each region.

^cThe total area of uncultivated soils uncultivated soils was 739.71 and 137.71 Mha, respectively.

227 deficiency of water along with preferential removal of topsoil rich in SOC by
 228 erosion (Tisdall 1996). Relatively low topsoil SOC levels and reductions due to
 229 cultivation could be found in South China and East China where traditional
 230 agriculture was characterized by well managed practices for enhancing soil
 231 fertility for a long time (Li 1992; He 1994). A remarkable increase of topsoil
 232 SOC density has been observed in cropland soils in these regions shifting from
 233 triple cropping to double cropping since the 1980's (Pan et al. 2003b; Zhang
 234 et al. 2004).

235 *C stock of topsoil and cultivation-induced change in China*

236 An estimate of topsoil SOC and pool reduction of soils after cultivation was
 237 conducted by using the different statistical mean values obtained in this work
 238 (Table 3). The mean reduction of the SOC varied from 13 to 15 t Cha⁻¹ and
 239 the calculated pool reduction thus may lie between 1.7 to 2.0 Pg. This was
 240 apparently smaller than the sum of the pool reduction for all the geographical
 241 regions. Error may exist because sampling intensity in some regions (such as
 242 the Tibet plateau and Northwest China) was not sufficient as compared to the
 243 eastern China (cf: Figure 1). There is still uncertainty in estimating the overall
 244 pool of topsoil SOC of uncultivated soils as a large portion of them were not
 245 surveyed. Overall topsoil SOC pool of uncultivated soils of China could be
 246 estimated amounting to 36.86 Pg by using the mean SOC density of 50 t Cha⁻¹
 247 for the surveyed uncultivated soils (Table 1). When taking into account that
 248 the unsurveyed soils were mainly in the Tibet Plateau (Figure 1) due to inac-
 249 cessibility to soil sampling in the early 1980's, the rest uncultivated soils may
 250 have a total topsoil C pool of 12.82 Pg using a mean topsoil SOC density of
 251 42 t Cha⁻¹ for China alpine soils. Therefore, a reasonable overall pool of the
 252 uncultivated soils of China may be 40.4–42.0 Pg. Nevertheless, long-term
 253 cultivation of China had induced a reduction of topsoil SOC around
 254 14 t Cha⁻¹ and an overall pool loss of 2 Pg after cultivation of natural soils.
 255 These C losses were especially remarkable in Northeast and North China.

256 Wu et al. (2003) made an estimation of total SOC stock of China's soils
 257 being 78.3 Pg in upper 1 m by using similar methodology and recently Li
 258 reported an estimate at 82.6 Pg by using a biogeochemical model. Accordingly,

Table 3. Estimate of loss of topsoil SOC density and pool of soils after cultivation.

Soils	Mean SOC density (t Cha ⁻¹)			C pool (Pg)	
	Soil profile statistics	Soil region statistics	Weighed by soil area	Surveyed area	Total area
Uncultivated	49.84 ± 46.69	51.02 ± 66.40	51.98	22.50	35.25
Cultivated soils	35.08 ± 31.57	38.41 ± 31.15	36.97	4.81	5.12
Cultivation-induced loss	14.76	12.61	15.01	1.63–1.94	1.74–2.07

259 the topsoil SOC pool size here accounted for 48–54% of the total of their
 260 estimates. The total topsoil SOC pool of China amounted to 5.1–5.3% to the
 261 world total in contrast to the soil area proportion of 6.4% of the world. In
 262 contrast, the soil area of French was 0.30% of the world and the total topsoil
 263 SOC storage was 0.47% of world's total (Arrouys and Balesdent 2002), and
 264 Brazil's area was 5.54% of the world, with the 5.32% of total topsoil SOC
 265 storage (Bernoux et al. 2002). Thus, China could be considered as a country of
 266 low topsoil C. The total cultivation-induced loss 2 Pg of topsoil SOC consti-
 267 tuted 29% of the total cultivated C loss of China soils in 1 m depth of soil (Wu
 268 et al. 2003) and as much as 40% of the present stock of the cultivated soils of
 269 China.

270 While China is still facing the challenge of soil degradation, C sequestration
 271 is important for China under Kyoto Protocol. The present low topsoil SOC
 272 density may offer potential for C sequestration in agriculture when adopting C
 273 sequestration strategy and practical measures (Lal 2002b). The C loss could be
 274 expected to recoverable by conservation tillage, along with efficient manage-
 275 ment of inputs of irrigation, fertilizer, and pesticides in agricultural systems.
 276 Several authors have shown considerable rate of topsoil SOC increase in China
 277 agricultural soils (SESATC 2003), particularly in paddy soils (Pan et al. 2003a,
 278 b; Zhang et al. 2004). However, approaches for attaining rapid topsoil SOC
 279 sequestration deserve research needs for China soil studies on soil C policies
 280 and national C sequestration strategy.

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284 References

- 285 Amundson R. 2001. The carbon budget in soils. *Annu. Rev. Earth Planet. Sci.* 29: 535–562.
 286 Anonymous 2003. Suggestions for integrated control of land desertification in karst terrains of
 287 Southwest China. *Adv. Earth Sci.* 18: 489–492. (in Chinese).
 288 Arrouys D. and Balesdent J. 2002. Increasing Carbon Stocks in French Agricultural Soils. Scientific
 289 Assessment Unit for Expertise, INRA, <http://www.inra.fr/actualites/rapport-carbone.html>.
 290 Batjes N.H. 1996. Total carbon and nitrogen in the soils of the world. *Eur. J. Soil Sci.* 47: 151–163.
 291 Batjes N.H. 2002. Carbon and nitrogen stocks in the soils of Central and Eastern Europe. *Soil Use*
 292 *Manage.* 18: 324–329.
 293 Batjes N.H. and Dijkshoorn J.A. 1999. Carbon and nitrogen stocks in the soils of the Amazon
 294 region. *Geoderma* 89: 273–286.
 295 Bernoux M., Carvalho M.D.S and Volkoff B. et al. 2002. Brazil's soil carbon stocks. *Soil Sci. Soc.*
 296 *Am. J.* 66: 888–896.
 297 Bhattacharya T., Pal D.K. and Mandal C. et al. 2000. Organic carbon stock in Indian soils and
 298 their geographical distribution. *Curr. Sci.* 79: 655–660.
 299 Bhatti J.S., Apps M.J. and Jiang H. 2002a. Influence of nutrients, disturbances and site conditions
 300 on carbon stocks along a boreal forest transect in central Canada. *Plant Soil* 242: 1–14.

- 301 Bhatti J.S., Apps M.J. and Tarnocai C. 2002b. Estimates of soil organic carbon stocks in central
302 Canada using three different approaches. *Can. J. Forest Res.* 32: 805–812.
- 303 Bousquet P., Ciais P. and Peylin P. et al. 1999. Inverse modeling of annual atmospheric CO₂
304 sources and sinks I. Method and control inversion. *J. Geophys. Res.* 104: 26161–26178.
- 305 Callesen I., Liski J. and Raulund-Rasmussen K. et al. 2003. Soil carbon stores in Nordic well-
306 drained forest soils-relationships with climate and texture class. *Glob. Change Biol.* 9: 358–370.
- 307 Dadal R.C. and Mayer R.J. 1986. Long-term trends in fertility of soils under continuous cultivation
308 and cereal cropping in southern Queensland. II. Total organic carbon and its rate of loss from
309 the soil profile. *Aust. J. Soil Res.* 24: 281–291.
- 310 Department of Agriculture, Hunan Province 1989. *Soils of Hunan*. China Agriculture Press,
311 Beijing(in Chinese).
- 312 Esser G. 1987. Sensitivity of global carbon Pools and fluxes to human and potential climatic
313 impacts. *Tellus* 39B: 245–260.
- 314 Eve M.D., Sperow M. and Howerton K. et al. 2002. Predicted impact of management changes on
315 soil carbon storage for each cropland region of the conterminous United States. *J. of Soil Water*
316 *Conserv.* 57: 196–204.
- 317 FAO & UNESCO 1988. *Soil Map of the World. Revised Legend*, Rome.
- 318 Fearnside P.M. and Barbosa R.I. 1998. Soil carbon changes from conversion of forest to pasture in
319 Brazilian Amazonia. *Forest Ecol. Manage.* 108: 117–166.
- 320 Hao Y., Lal R. and Owens L.B. et al. 2002. Effect of cropland management and slope position on
321 soil organic carbon pool at the North Appalachian Experimental Watersheds. *Soil Till. Res.* 68:
322 133–142.
- 323 He D.Y. 1994. *Fertility of soils in South China and Crop Fertilization*. Science Press, Beijing,
324 China, pp.19–27(in Chinese).
- 325 Houghton R.A., Hackle J.L. and Lawrence K.T. et al. 1999. The US carbon budget: contributions
326 from land-use change. *Science* 285: 574–578.
- 327 Jacinthe P.A., Lal R. and Kimble J.M. 2001. Organic carbon storage and dynamics in croplands
328 and terrestrial deposits as influenced by subsurface tile drainage. *Soil Sci.* 166: 322–335.
- 329 Kirschbaum M.U.F. 2000. Will changes in soil organic carbon act as a positive or negative feed-
330 back on global warming? *Biogeochemistry* 48: 21–51.
- 331 Lal R. 1999. World soils and greenhouse effect. *IGBP Glob. Change Newslett.* 37: 4–5.
- 332 Lal R. 2002a. Carbon sequestration in dryland ecosystems of West Asia and North Africa. *Land*
333 *Degrad. Dev.* 13: 45–59.
- 334 Lal R. 2002b. Soil carbon sequestration in China through agricultural intensification, and resto-
335 ration of degraded and desertified ecosystems. *Land Degrad. Dev.* 13: 469–478.
- 336 Lal R., Follett R.F. and Kimble J. et al. 1999. Managing US cropland to sequester carbon in soil.
337 *J. Soil Water Conserv.* 55: 374–381.
- 338 Li Q. 1992. *Paddy Soils of China*. China Science Press, Beijing, China, pp. 232–248(in Chinese).
- 339 Li C. 2000. Decrease of soil organic carbon pool: risk of China's agriculture. Comparison of C
340 cyclings in agro-ecosystems between China and US. *Quaternary Sci.* 20: 345–350(in Chinese).
- 341 Li K., Wang S.Q. and Cao M.K. 2004. Vegetation and soil C storage in China. *Sci. China, Ser. D*
342 47: 49–57.
- 343 Li Z. and Zhao Q.G. 2001. Organic carbon content and distribution in soils under different land
344 uses in tropical and subtropical China. *Plant Soil* 231: 175–185.
- 345 Lindert P.H., Lu J. and Wu W. 1996. Trends in the soil chemistry of South China since the 1930s.
346 *Soil Sci.* 161: 329–342.
- 347 Liu X., Lai Q. and Huang Q. 1999. Soil organic matter dynamics in red soils region of Jiangxi.
348 *Jiangxi J. Agr. Sci.* 11(Supplement): 14–23.
- 349 Ni J. 2001. Carbon storage in terrestrial ecosystem of China: estimates at different spatial
350 resolutions and response to climatic change. *Climatic Change* 49: 339–358.
- 351 Pan G., Li L., Zhang X. and Wu L. 2003a. Storage and sequestration potential of topsoil organic
352 carbon in China's paddy soils. *Glob. Change Biol.* 10: 79–92.

- 353 Pan G., Li L. and Zhang X. et al. 2003b. Soil organic carbon storage of China and the seques-
354 tration dynamics in agricultural lands. *Adv. Earth Sci.* 18: 609–618(in Chinese).
- 355 Rounsevell M.D.A, Evans S.P. and Bullock P. 1999. Climate change and agricultural soils: impacts
356 and adaptation. *Climate Change* 43: 683–709.
- 357 Rustad L.E., Campbell J.L. and Marion G.M. et al. 2001. Meta-analysis of the response of soil
358 respiration, net nitrogen mineralization, and aboveground plant growth to experimental
359 ecosystem warming. *Oecologia* 26: 543–562.
- 360 Schlesinger W.H. 1999. Carbon sequestration in soils. *Science* 284: 2095.
- 361 Schlesinger W.H. 2000. Response. *Science* 288: 811.
- 362 Schuman G.E., Janzen H.H. and Herrick J.E. 2002. Soil carbon dynamics and potential carbon
363 sequestration by rangelands. *Environ. Pollut.* 116: 391–396.
- 364 Smith P., Milne R. and Powlson D.S. et al. 2000a. Revised estimates of the carbon mitigation
365 potential of UK agricultural land. *Soil Use Manage.* 16: 293–295.
- 366 Smith P., Powlson D.S. and Smith J.U. et al. 2000b. Meeting Europe's climate change commit-
367 ments: quantitative estimates of the potential for carbon mitigation by agriculture. *Glob. Change*
368 *Biol.* 6: 525–539.
- 369 State Extension Service of Agricultural Technology of China (SESATC) 2003. State-wide Soil
370 Monitoring of Arable Lands (Selected Works). China Agriculture Press, Beijing(in Chinese).
- 371 State Soil Survey Service of China (SSSSC) 1993. China Soil Types, Vol. 1. China Agricultural
372 Press, Beijing(in Chinese).
- 373 SSSSC 1994a. China Soil Types, Vol. 2. China Agricultural Press, Beijing(in Chinese).
- 374 SSSSC 1994b. China Soil Types, Vol. 3. China Agricultural Press, Beijing(in Chinese).
- 375 SSSSC 1995a. China Soil Types, Vol. 4. China Agricultural Press, Beijing(in Chinese).
- 376 SSSSC 1995b. China Soil Types, Vol. 5. China Agricultural Press, Beijing(in Chinese).
- 377 SSSSC 1996a. China Soil Types, Vol. 6. China Agricultural Press, Beijing(in Chinese).
- 378 SSSSC 1996b. Soil Survey Technical Report. China Agriculture Press, Beijing.
- 379 SSSSC 1997. China Soil Survey Database. China Agricultural Press, Beijing(in Chinese).
- 380 SSSSC 1998. China Soils. China Agricultural Press, Beijing(in Chinese).
- 381 Tian H., Hall C.A.S. and Qi Y. 2000. Increased biotic metabolism of the biosphere inferred from
382 observed data and model. *Sci. China, Types B (Chemistry)* 40: 58–68.
- 383 Tisdall J.M. 1996. Formation of soil aggregates and accumulation of soil organic matter. In: Carter
384 M.R. and Stewart B.A. (eds), *Structure and Organic Matter Storage Agricultural Soils.*
385 *Advances in Soil Science.* CRC Press/Lewis Publishers, New York, pp. 57–96.
- 386 Uri N.D. 2001. The potential impact of conservation practices in US agriculture on global climate
387 change. *J. Sustain. Agr.* 18: 109–131.
- 388 Vleeshouwers L.M. and Verhagen A. 2002. Carbon emission and sequestration by agricultural land
389 use: a model study for Europe. *Glob. Change Biol.* 8: 519–530.
- 390 Wang J., Fu B.J. and Qiu Y. et al. 2001. Soil nutrients in relation to land use and landscape
391 position in the semi-arid small catchment on the Loess Plateau in China. *J. Arid Environ.* 48:
392 537–550.
- 393 Wang J.Z., Ma Y. and Jin G. 1996. China Irragric Soils. Science Press, Beijing(in Chinese).
- 394 Wang S.Q. and Zhou C.H. 1999. Estimate of organic carbon pool of terrestrial soils of China.
395 *Geogr. Sci.* 18: 349–355(in Chinese).
- 396 Wang S.Q., Zhou C.H. and Liu J.Y. et al. 2002. Carbon storage in northeast China as estimated
397 from vegetation and soil inventories. *Environ. Pollut.* 116((Suppl.): 157–165.
- 398 West T.O. and Marland G. 2002. A synthesis of carbon sequestration, carbon emissions, and net
399 carbon flux in agriculture: comparing tillage practices in the United States. *Agr. Ecosyst. Envi-*
400 *ron.* 91: 217–232.
- 401 Wu H.B., Guo Z.T. and Peng C.H. 2003. Land use induced changes of organic carbon storage in
402 soils of China. *Glob. Change Biol.* 9: 305–315.
- 403 Xie Y. 1999. Analysis of sensitivity of China food production in response to climatic change. *Sci.*
404 *Resour.* 21: 13–18(in Chinese).

- 405 Zhang Q., Pan G. and Li L. et al. 2004. Dynamics of topsoil organic carbon of paddy soils from at
406 Yixing over the last 20 years and the driving factors. *Quaternary Sci.* 24: 114–120(in Chinese).
- 407 Zhao Q.G., Zhang L. and Xia Y.F. 1997. Organic carbon storage in soils of Southeast China. *Nutr.*
408 *Cycl. Agroecosys.* 49: 229–234.
- 409 Zheng Y., Zhou G. and Zhang X. et al. 1997. Sensitivity of terrestrial ecosystems to global change
410 in China. *Acta Bot. Sin.* 39: 837–840(in Chinese).
- 411 Zhou G. 1999. Impact of climate change on NPP of agriculture and animal husbandry in
412 ecologically vulnerable areas: mechanism and modeling. *Resour. Sci.* 21: 46–50(in Chinese).
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